

Increases in Greenhouse Gases Following the Use of Peatlands for Agricultural Areas

by Noor Aidawati

Submission date: 13-Aug-2024 11:39PM (UTC-0400)

Submission ID: 2431806596

File name: Gases_Following_the_Use_of_Peatlands_for_Agricultural_Areas.pdf (940.81K)

Word count: 4127

Character count: 22139

Increases in Greenhouse Gases Following the Use of Peatlands for Agricultural Areas

Akhmad R. Saidy^{1,2*}, Fakhrr Razie¹, Noor Aidawati¹, Taufik Hidayat¹

¹ Department of Soil, Faculty of Agriculture, Lambung Mangkurat University, Jalan Jenderal A. Yani KM 36, Simpang Empat Banjarbaru, Kalimantan Selatan 70714, Indonesia.

² Doctoral Program of Agricultural Science, Postgraduate Program, Lambung Mangkurat University, Jalan Jenderal A. Yani KM 36, Simpang Empat Banjarbaru, Kalimantan Selatan 70714, Indonesia.

* Corresponding Author: asaidy@ulm.ac.id

Abstract. Peatlands with very high C contents are generally considered a source of greenhouse gas (GHG) emissions. This research aimed to quantify the changes in peatland characteristics and GHG emission from the conversion of peatlands to agricultural areas. Height of water table, pH, electrical conductivity (EC), redox potential (E_h), organic carbon (OC), hot water-soluble C, total-N, the concentrations of NH_4^+ and NO_3^- , soluble-Al, soluble-Fe, and the emission of CH_4 , CO_2 and N_2O were quantified before and after land clearing of peats for agriculture. Results of study showed that pH, EC, OC contents, hot water-soluble-C, and total-N did not change after peatland clearing for agriculture. On the other hand, the concentrations of NH_4^+ , NO_3^- , soluble-Al, soluble-Fe, redox potential and height of water table increased significantly after the peatland land clearing. Methane emissions from peatlands before peatland clearing were in the range of 0.13-0.22 $\text{mg C m}^{-2} \text{ h}^{-1}$ increased significantly to 0.14-0.31 $\text{mg C m}^{-2} \text{ h}^{-1}$ after the peatland clearing. The land clearing of peatlands for agricultural practices also caused increases in CO_2 and NO_x emissions by 85% and 76%, respectively. Changes in GHG emissions were not related to the changes in substrates quality of peats (OC contents, total-N and hot water-soluble-C). Results of the study indicate that increases in the GHG emission following the use of peatlands for agricultural areas are attributed to the changes in the peat characteristics.

1. Introduction

Peatlands in Indonesia, which are commonly found in Kalimantan, Sumatra, and Papua, are part of world's tropical peatlands and estimated to range from 16.8 to 27.0 million ha [1]. Peatlands with high rates of organic matter accumulation and low rate of decomposition are considered as sink of C. The amount of C stored in Indonesian peatlands is estimated to reach 55 Gt [2]. Peatlands are also capable of depositing water up to 10 times their mass [3,4]; therefore, peatlands are able to store large amounts of water during the rainy season and will then release slowly during the dry season [5,6]. Results of these studies indicate that peatlands in Indonesia play an important role in environments.

Some of Indonesia's peatlands have been reclaimed and used for agricultural lands in the last two decades, especially for large-scale oil palm plantation [7]. Conversion of peatlands to agricultural



Content from this work may be used under the terms of the [Creative Commons Attribution 3.0 licence](https://creativecommons.org/licenses/by/3.0/). Any further distribution of this work must maintain attribution to the author(s) and the title of the work, journal citation and DOI.

lands causes an increase in GHG emissions from reclaimed peatlands [8,9]. The use of 416,000 ha of deep peatlands for oil palm plantations in the Ex Mega Rice Project of Central Kalimantan, Indonesia is projected for the next 25 years to emit 93–217 Mt CO₂ equivalent [10]. The conversion of peatlands into agricultural lands may cause Indonesia to be a contributor country to GHG emissions in the world.

Information on the mechanism of increasing GHG emissions due to the conversion of peatlands to agricultural areas is very crucial for the GHG emission mitigation of reclaimed-peatlands. Results of previous studies showed that long-term conversion of peatlands for agriculture results in changes in substrate quality of peats [11,12], which eventually results in changes in the GRG emission of peatlands. The use of tropical peatlands for paddy cultivation for 15 years results in a decrease in carbohydrate content and an increase in lignin content [13]. Peat with O-alkyl-C structures decrease significantly, while peat alkyl-C structures show a significant increase after a-20-year agricultural practices in tropical peatlands [14]. Results of these studies imply that changes in the GHG emission due to long-term use of peatlands for agricultural is related changes in the substrate quality of peats. However, comprehensive information on factors controlling changes in the GHG emission due to short-term conversion of peatlands to agricultural areas is unavailable. Therefore, the objective of this study was to quantify the immediate changes in the emissions of CH₄, CO₂, and N₂O following the use of peatlands for agricultural areas.

2. Materials and Methods

2.1. Experimental site and the measurement of gases

The study site is administratively situated in the Desa Landasan Ulin Utara (3°09'56.6"–3°10'41.0" S; 114°37'58.3"–114°39'51.4"E), Sub-District of Liang Anggang, Kota Banjarbaru, Province of South Kalimantan. The average temperature in the areas was 27-28 °C, and the annual precipitation was 2,600 mm. Dominant vegetation in this peatlands were *Melaleuca cajuputi* Roxb, *Melastoma malabathricum* L., and *Stenochlaena palustris* (Burm.F) Bedd.

Sampling and measurements of greenhouse gas emissions and peat characteristics were carried out at 15 different sampling points twice, before the land clearing (last week of November 2017) and after land clearing (second week of December 2017). Gas sampling was carried out using a static chamber method. The gas was collected at interval of 4, 8, 12 and 15 minutes after the closure of chamber using a 10 mL syringe through the septum in the middle of the chamber, then the collected gas was transferred to a 10 mL air-tight glass vial. The gas was then injected onto the Shimadzu GC-14A equipped with electron capture detector for N₂O measurement, flame ionization detector for CH₄ measurement, and thermal conductivity detector for CO₂ measurement. The calculation of gas fluxes were carried out using linear regressions of the change in gas concentration, the volume of chamber, and soil surface area and corrected for the field measured air temperature and atmospheric pressure [15].

2.2. Peat sampling and characterization

Peat samples were collected from each sampling point at a depth of 30 cm. Soil pH, electrical conductivity, height of water level, and redox potential were measured directly in the field. After removing plant debris, the samples were stored at 4 °C until the determination of peat chemical characteristics. Ammonium and NO₃ were extracted from peats using KCl 1 N (1:10=weight:volume), and the concentrations of NH₄ and NO₃ in the extract were measured using a hydrazine reduction method for NO₃-N and an indophenol blue method for NH₄-N [16]. The amount of organic-C in the peat was determined using the wet oxidation method of Walkley-Black [17], and total-N in the peats was measured using the Kjeldahl method [18]. Hot water soluble-C was extracted from the peat using warm water (60 °C), and the concentration of organic-C in the extract was quantified using the anthrone-sulfuric acid method [19]. Soluble-Al and -Fe were extracted from the peat using NH₄OAc 1 N pH 4.8, and the concentrations of Al and Fe in the extract were quantified using aluminon and o-penanthroline methods, respectively [20,21].

2.3. Statistical analysis

Comparison of mean using t-test was conducted to assess if there are differences in the means of variables before and after land clearing of peats for agriculture. Relationship between the emission of CH₄, CO₂, N₂O and soil characteristics of peats was measured using correlation-regression analyses. All statistical analyses were performed using GENSTAT 12th Edition [22].

3. Results and Discussion

The conversion of peatlands to agriculture areas resulted in an immediate increases in the emissions of CH₄, CO₂, and N₂O (Figure 1). Methane emission before peat land clearing for agriculture was higher than that before land clearing ($P < 0.05$). Methane emission before land clearing ranged from 0.13 to 0.22 mg C m⁻² h⁻¹ (average 0.17 mg C m⁻² h⁻¹), while range of 0.13–0.31 mg C m⁻² h⁻¹ (average 0.23 mg C m⁻² h⁻¹) CH₄ emission was observed after land clearing (Figure 1). Land clearing of peatlands for agriculture also resulted significant increases in CO₂ and N₂O emission by 85.48% and 76.34%, respectively ($P < 0.05$).

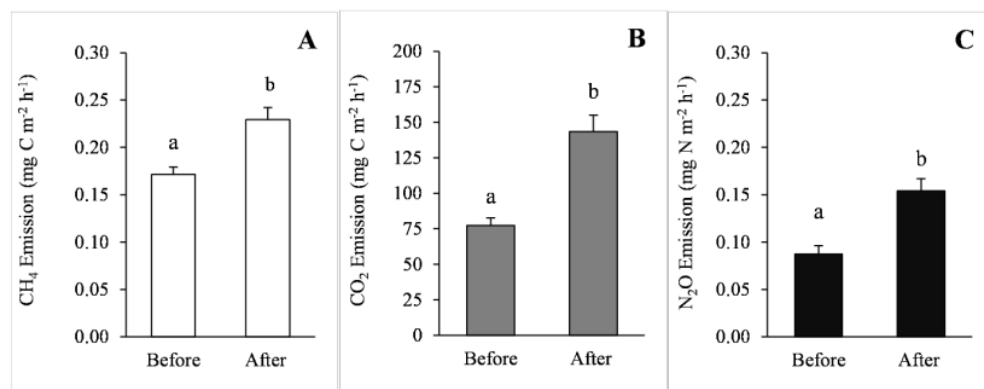


Figure 1: Emissions of CH₄ (A), CO₂ (B), and N₂O (C) before and after land clearing of peats for agriculture. The vertical bars represent standard errors of mean (n=15). Similar letters above columns indicate no statistical differences in GHG emissions between before and after land clearing of peats for agriculture based on the mean difference test at $P < 0.05$.

It well known that the GHG emission are significantly influenced by environmental factors such as soil pH, redox potential, temperature [23,24] and changes in substrate quality of organic-C (peats) [25,26]. Peat characteristics before and after land clearing of peatlands were quantified to determine the relationship between changes in the GHG emissions and changes in peat characteristics before and after land clearing. Several peat characteristics such as peat pH, EC, organic-C and total-N contents, and hot water-soluble-C content were not significantly different between before and after land clearing ($P > 0.05$; Table 1). On the other hand, land clearing of peatlands for agriculture led to considerable increases in the height of water table, redox potential, the concentrations of NH₄ and NO₃, and the concentration of soluble -Al and -Fe ($P < 0.05$; Table 1).

No.	Peat Characteristics	Before	After
1.	Peat pH (H ₂ O, 1:5)	3.49 ± 0.05* a**	3.52 ± 0.05 a
2.	Electrical Conductivity (mS)	391.99 ± 9.07 a	368.18 ± 6.25 a
3.	NH ₄ ⁺ (mg N kg ⁻¹)	9.76 ± 0.94 a	20.10 ± 2.14 b
4.	NO ₃ ⁻ (mg N kg ⁻¹)	9.94 ± 1.94 a	22.16 ± 1.37 b
5.	Height of water table (cm)	2.11 ± 3.89 a	-8.56 ± 2.24 b
6.	Redox potential – E _h (mV)	-7.29 ± 6.81 a	-71.47 ± 9.87 b
7.	Organic C (g C kg ⁻¹)	189.59 ± 6.59 a	192.56 ± 6.48 a
8.	Total N (g N kg ⁻¹)	12.84 ± 0.92 a	12.55 ± 0.92 a
9.	Hot water soluble C (mg C kg ⁻¹)	32.43 ± 5.29 a	31.18 ± 5.30 a
10.	Soluble Al (mg Al kg ⁻¹)	11.38 ± 1.05 a	22.64 ± 1.90 b
11.	Soluble Fe (mg Fe kg ⁻¹)	15.37 ± 1.62 a	32.93 ± 2.73 b

* Standard error of mean (n=15). ** Different small case indicate statistically differences between peat characteristics before and after land clearing of peats for agriculture based on the mean difference test at $P < 0.05$.

Table 1: Changes in peat characteristics before and after the use of peatlands for agriculture.

Results of the previous studies stated that changes in the emission of GHG due to land changes are attributed the changes in the substrate quality [27,28]. However, the parameters of peat substrate quality in this study such as organic-C, total-N, and hot water soluble-C were not significantly different between before and after land clearing of peatland for agriculture ($P > 0.01$; Table 1) and were not related to the changes in the GHG emission (data not shown). No significant changes in the parameters of peat substrate quality and no significant relation between substrate quality of peats and the GHG emission are probably due to short period before after land clearing (only three weeks) was not able to change the substrate quality of peats. Therefore, changes in the GHG emission in this study was likely to be attributed to the changes in the chemical characteristics of peats before and after land clearing. Results of the analyses of correlation and regression showed that the emission of CH₄ was related significantly to the changes in the height of water table and redox potential (Figure 2A and 2B), the emission of CO₂ was negatively correlated with the concentration of soluble-Al and -Fe (Figure 2C and 2D), and the emission of N₂O have significant correlation with the concentration of NO₃ and NH₄ (Figure 2E and 2F).

It was reported in the previous study that water table level control the emission of CH₄ from peatlands in China, in which the water level surface of peatlands drawdown from 0 to 50 cm below soil surface lead to 82% reduction in the CH₄ emission [29]. Results of study of tropical peat swamp forest also showed that increasing ground water level led to the development of anaerobic conditions, which is eventually stimulated CH₄ production [30]. Redox potential (E_h) is commonly used to determine the methanogenic activity. The methanogenic activity is negatively correlated with the E_h values, in which E_h of below +240 mV are reported to be threshold levels for methanogenic activity [31]. Relationship between redox potential and the emission of CH₄ in this study was in agreement with previous studies in the tropical peatlands [32,33].

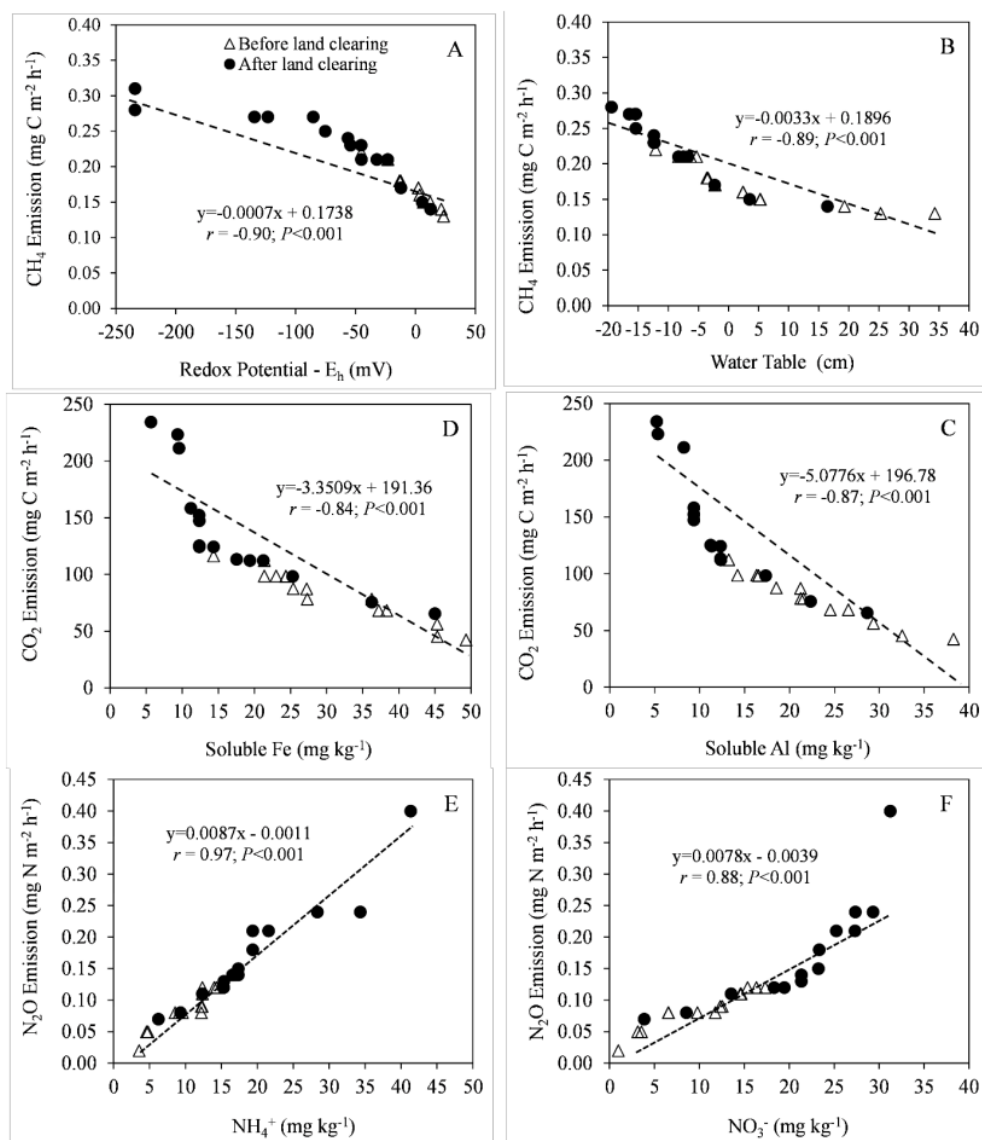


Figure 2. Relationship between CH₄ emissions and redox potential (A), CH₄ emissions and height of water table (B), CO₂ emissions and soluble-Al (C), CO₂ emissions and soluble-Fe (D), N₂O emissions and NH (E), and N₂O emissions and NO₃ (F).

Carbon dioxide emission of peatlands was in the range of 42.35–112.35 mg C m⁻² h⁻¹ before land clearing increased to 98.45–234.40 mg C m⁻² h⁻¹ after land clearing (Figure 2). The presence of Fe and Al in soils reduces C mineralization due the stabilization of organic-C through cation bridging and ligand exchange of organic-C by Al and Fe, which is eventually increasing the amount of organic-C protected from microbial decomposition [34,35]. Results of this study showed that the emission of

CO₂ was significantly and negatively correlated with soluble-Al and -Fe (Figure 2), indicating important roles of soluble Al and Fe in reducing C mineralization of peats. Results of this study are consistent with the previous studies stated that the presence Al and Fe decreases C mineralization [36,37].

The amounts of NH₄ and NO₃ are often considered to be the factors controlling the rate of N₂O emission through nitrification and denitrification processes [38,39]. The concentration of NH₄ and NO₃ increased from 3.56–15.25 mg N kg⁻¹ before land clearing to 12.35–41.32 mg N kg⁻¹ after land clearing and from 0.98–21.35 mg N kg⁻¹ before land clearing to 14.56–31.24 mg N kg⁻¹ after land clearing, respectively (Table 1 and Figure 2). Increasing the amounts of NH₄ and NO₃ following land clearing of peats for agriculture results in increasing the emission of N₂O. This was supported the significant and positive correlation between the emission of N₂O and the concentration of NH₄ and NO₃ (Figure 2).

4. Conclusion

The emissions of CH₄, CO₂, and N₂O increase significantly following land clearing of peatlands for agriculture. Land clearing of peatland for agriculture also result in significant changes in chemical characteristics peats, i.e. height of water table, redox potential (Eh), the concentration of NH₄ and NO₃, and the concentration of soluble-Al and -Fe. However, peat pH, electrical conductivity and the parameters of peat substrate quality such as organic-C, total-N, and hot water-soluble-C did not change following land clearing of peatlands for agriculture. Short period of peatland conversion for agriculture (three weeks) was thought to be unable to changes the substrate quality of peats. Results obtained in this study demonstrate that changes in the emission of CH₄, CO₂, and N₂O as a results of a short period of peatland conversion for agriculture were attributed to the changes in the chemical characteristics of peats.

Acknowledgments

This project was funded by the Peatland Restoration Agency of Indonesia through the Grant of Paludiculture Research Project 2017.

References

- [1]. Page, S.E. and Bank, C. (2007) *Tropical Peatlands: Distribution, Extent and Carbon Storage - Uncertainties and Knowledge Gap*. Peatland International.
- [2]. Jaenicke, J., Rieley, J.O., Mott, C., Kimman, P. and Siegert, F. (2008) Determination of the amount of carbon stored in Indonesian peatlands. *Geoderma*, **147**, 151-158.
- [3]. Winarna, K.M., Sabiham, S., Sutandi, A. and Sutarta, E.S. (2016) Hydrophobicity of Tropical Peat Soil from an Oil Palm Plantation in North Sumatra. *Journal of Agronomy*, **15**, 114-121.
- [4]. Taufik, M., Veldhuizen, A.A., Wösten, J.H.M. and van Lanen, H.A.J. (2019) Exploration of the importance of physical properties of Indonesian peatlands to assess critical groundwater table depths, associated drought and fire hazard. *Geoderma*, **347**, 160-169.
- [5]. Waddington, J., Kellner, E., Strack, M. and Price, J. (2010) Differential peat deformation, compressibility, and water storage between peatland microforms: Implications for ecosystem function and development. *Water Resources Research*, **46**.
- [6]. Bourgault, M.A., Larocque, M. and Garneau, M. (2017) Quantification of peatland water storage capacity using the water table fluctuation method. *Hydrological processes*, **31**, 1184-1195.
- [7]. Afriyanti, D., Kroeze, C. and Saad, A. (2016) Indonesia palm oil production without deforestation and peat conversion by 2050. *Science of the Total Environment*, **557**, 562-570.

- [8]. Wijedasa, L.S., Sloan, S., Page, S.E., Clements, G.R., Lupascu, M. and Evans, T.A. (2018) Carbon emissions from South-East Asian peatlands will increase despite emission-reduction schemes. *Glob. Change Biol.*, **24**, 4598-4613.
- [9]. Surahman, A., Soni, P. and Shivakoti, G.P. (2018) Reducing CO₂ emissions and supporting food security in Central Kalimantan, Indonesia, with improved peatland management. *Land Use Policy*, **72**, 325-332.
- [10]. Dohong, A., Aziz, A.A. and Dargusch, P. (2018) Carbon emissions from oil palm development on deep peat soil in Central Kalimantan Indonesia. *Anthropocene*, **22**, 31-39.
- [11]. Cooper, H.V., Vane, C.H., Evers, S., Aplin, P., Girkin, N.T. and Sjögersten, S. (2019) From peat swamp forest to oil palm plantations: The stability of tropical peatland carbon. *Geoderma*, **342**, 109-117.
- [12]. Könönen, M., Jauhiainen, J., Straková, P., Heinonsalo, J., Laiho, R., Kusin, K., Limin, S. and Vasander, H. (2018) Deforested and drained tropical peatland sites show poorer peat substrate quality and lower microbial biomass and activity than unmanaged swamp forest. *Soil Biology and Biochemistry*, **123**, 229-241.
- [13]. Saïdy, A.R., Mariana, Z.T., Adji, F.A., Nusantara, R.W., Fitria, I. and Syahrinuddin. (2018) Carbon mineralization dynamics of tropical peats in relation to peat characteristics. *Biodiversitas Journal of Biological Diversity*, **19**, 1413-1421.
- [14]. Saïdy, A.R. (2005) Struktur karbon organik yang ditetapkan dengan ¹³C Nuclear Magnetic Resonance (NMR) Spektroskopi dan mineralisasi karbon pada gambut yang digunakan untuk pertanian. *Jurnal Tanah Tropika*, **1**, 15-23.
- [15]. Sackett, T.E., Basiliko, N., Noyce, G.L., Winsborough, C., Schurman, J., Ikeda, C. and Thomas, S.C. (2015) Soil and greenhouse gas responses to biochar additions in a temperate hardwood forest. *GCB Bioenergy*, **7**, 1062-1074.
- [16]. Bundy, L.G. and Meisinger, J.J. (1994) In Weaver, R. W., Angle, J. S. and Bottomley, P. S. (eds.), *Methods of Soil Analysis, Part 2: Chemical and Biological Methods*. Soil Science Society of America, Madison WI, pp. 951-984.
- [17]. Nelson, D.W. and Sommers, L.E. (1996) In Sparks, D. L. (ed.), *Methods of Soil Analysis Part 3: Chemical Methods*. Soil Science Society of America-American Society of Agronomy Inc., Madison WI, pp. 961-1011.
- [18]. Bremer, J.M. and Mulvaney, C.S. (1982) In Page, A. L. and Keeney, D. R. (eds.), *Methods of Soil Analysis Part 2: Chemical and Biological Methods*. Soil Science Society of America Inc., Madison WI, pp. 595-709.
- [19]. Grandy, A.S., Erich, M.S. and Porter, G.A. (2000) Suitability of the anthrone-sulfuric acid reagent for determining water soluble carbohydrates in soil water extracts. *Soil Biology & Biochemistry*, **32**, 725-727.
- [20]. Fortune, W. and Mellon, M. (1938) Determination of iron with o-phenanthroline: a spectrophotometric study. *Industrial & Engineering Chemistry Analytical Edition*, **10**, 60-64.
- [21]. Hsu, P.H. (1963) Effect of initial pH, phosphate, and silicate on the determination of aluminum with aluminon. *Soil Science*, **96**, 230-238.
- [22]. Payne, R. (2008) *A Guide to Anova and Design in Genstat*. VSN International, Hempstead, UK.
- [23]. Dowhower, S.L., Teague, W.R., Casey, K.D. and Daniel, R. (2020) Soil greenhouse gas emissions as impacted by soil moisture and temperature under continuous and holistic planned grazing in native tallgrass prairie. *Agriculture, Ecosystems & Environment*, **287**, 106647.
- [24]. Taft, H.E., Cross, P.A. and Jones, D.L. (2018) Efficacy of mitigation measures for reducing greenhouse gas emissions from intensively cultivated peatlands. *Soil Biology and Biochemistry*, **127**, 10-21.
- [25]. Padhy, S.R., Bhattacharyya, P., Dash, P.K., Reddy, C.S., Chakraborty, A. and Pathak, H. (2019) Seasonal fluctuation in three mode of greenhouse gases emission in relation to soil

- labile carbon pools in degraded mangrove, Sundarban, India. *Science of The Total Environment*, 135909.
- [26]. Hoyos-Santillan, J., Lomax, B.H., Large, D., Turner, B.L., Lopez, O.R., Boom, A., Sepulveda-Jauregui, A. and Sjögersten, S. (2019) Evaluation of vegetation communities, water table, and peat composition as drivers of greenhouse gas emissions in lowland tropical peatlands. *Science of The Total Environment*, **688**, 1193-1204.
- [27]. Dhandapani, S., Ritz, K., Evers, S., Yule, C.M. and Sjögersten, S. (2019) Are secondary forests second-rate? Comparing peatland greenhouse gas emissions, chemical and microbial community properties between primary and secondary forests in Peninsular Malaysia. *Science of The Total Environment*, **655**, 220-231.
- [28]. Sjögersten, S., Aplin, P., Gauci, V., Peacock, M., Siegenthaler, A. and Turner, B.L. (2018) Temperature response of ex-situ greenhouse gas emissions from tropical peatlands: Interactions between forest type and peat moisture conditions. *Geoderma*, **324**, 47-55.
- [29]. Yang, G., Chen, H., Wu, N., Tian, J., Peng, C., Zhu, Q., Zhu, D., He, Y., Zheng, Q. and Zhang, C. (2014) Effects of soil warming, rainfall reduction and water table level on CH₄ emissions from the Zoige peatland in China. *Soil Biology and Biochemistry*, **78**, 83-89.
- [30]. Ishikura, K., Hirata, R., Hirano, T., Okimoto, Y., Wong, G.X., Melling, L., Aeries, E.B., Kiew, F., San Lo, K. and Musin, K.K. (2019) Carbon dioxide and methane emissions from peat soil in an undrained tropical peat swamp forest. *Ecosystems*, **22**, 1852-1868.
- [31]. Szafranek-Nakonieczna, A. and Stepniewska, Z. (2015) The influence of the aeration status (ODR, Eh) of peat soils on their ability to produce methane. *Wetlands ecology and management*, **23**, 665-676.
- [32]. Tokarz, E. and Urban, D. (2015) Soil redox potential and its impact on microorganisms and plants of wetlands. *Journal of Ecological Engineering*, **16**, 20–30.
- [33]. Girkin, N.T., Turner, B.L., Ostle, N. and Sjögersten, S. (2018) Composition and concentration of root exudate analogues regulate greenhouse gas fluxes from tropical peat. *Soil Biology and Biochemistry*, **127**, 280-285.
- [34]. Chen, S., Hong, H., Huang, X., Fang, Q., Yin, K., Wang, C., Zhang, Y., Cheng, L. and Algeo, T.J. (2018) The role of organo-clay associations in limiting organic matter decay: Insights from the Dajiuhe peat soil, central China. *Geoderma*, **320**, 149-160.
- [35]. Ramesh, T., Bolan, N.S., Kirkham, M.B., Wijesekara, H., Kanchikerimath, M., Srinivasa Rao, C., Sandeep, S., Rinklebe, J., Ok, Y.S., Choudhury, B.U. *et al.* (2019) In Sparks, D. L. (ed.), *Advances in Agronomy*. Academic Press, Vol. 156, pp. 1-107.
- [36]. Saidy, A., Smernik, R., Baldock, J., Kaiser, K. and Sanderman, J. (2015) Microbial degradation of organic carbon sorbed to phyllosilicate clays with and without hydrous iron oxide coating. *European Journal of Soil Science*, **66**, 83-94.
- [37]. Van De Vreken, P., Gobin, A., Baken, S., Van Holm, L., Verhasselt, A., Smolders, E. and Merckx, R. (2016) Crop residue management and oxalate-extractable iron and aluminium explain long-term soil organic carbon sequestration and dynamics. *European Journal of Soil Science*, **67**, 332-340.
- [38]. Mueller, D., Bange, H.W., Warneke, T., Rixen, T., Mueller, M., Mujahid, A. and Notholt, J. (2016) Nitrous oxide and methane in two tropical estuaries in a peat-dominated region of northwestern Borneo. *Biogeosciences (BG)*, **13**, 2415-2428.
- [39]. Päm, J., Verhoeven, J.T., Butterbach-Bahl, K., Dise, N.B., Ullah, S., Aasa, A., Egorov, S., Espenberg, M., Järveoja, J. and Jauhiainen, J. (2018) Nitrogen-rich organic soils under warm well-drained conditions are global nitrous oxide emission hotspots. *Nature communications*, **9**, 1135.

Increases in Greenhouse Gases Following the Use of Peatlands for Agricultural Areas

ORIGINALITY REPORT

18%

SIMILARITY INDEX

14%

INTERNET SOURCES

14%

PUBLICATIONS

4%

STUDENT PAPERS

MATCH ALL SOURCES (ONLY SELECTED SOURCE PRINTED)

2%

★ www.repo.uni-hannover.de

Internet Source

Exclude quotes On

Exclude matches Off

Exclude bibliography On

Increases in Greenhouse Gases Following the Use of Peatlands for Agricultural Areas

GRADEMARK REPORT

FINAL GRADE

GENERAL COMMENTS

/0

PAGE 1

PAGE 2

PAGE 3

PAGE 4

PAGE 5

PAGE 6

PAGE 7

PAGE 8
